

Diel differences in the exploitation of shallow inshore habitats by three small bottom-dwelling fishes with implications for their monitoring in a large lowland river

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Abstract. Sampling small-bodied fish species in large lowland rivers poses numerous difficulties often resulting in the underestimation of their true distribution and abundance. Some of these drawbacks might be overcome by using additional sampling gears and adapting sampling strategies to the biological specifications of the target species. The results of a two-year, repeated survey on a single location on the Vistula River (southern Poland) revealed that nighttime beach seining proved to be very effective in collecting specimens of three imperiled species: *Romanogobio belingi*, *Romanogobio kesslerii*, and *Sabanejewia baltica*. It was found that these species were 4.3-8.8 times more abundant in nighttime than in daytime collections. The abundance of all three species peaked at dusk and gradually decreased in subsequent hours. In conclusion, nighttime beach seining is recommended as a supplementary sampling technique for small bottom-dwelling fish species in large lowland rivers.

Keywords: day vs. night sampling, Natura 2000, rare species, temporal variation, Vistula River

Introduction

Large rivers are especially difficult for monitoring small-bodied, bottom-dwelling fish species. Wide riverbeds, deep waters, swift currents, and the complexity of river systems often make sampling extremely inefficient (Casselman et al. 1990, De Leeuw et al. 2007, Lapointe et al. 2008, Freedman et al. 2009, Herzog et al. 2009, Szalóky et al. 2014, Zajicek and Wolter 2018). In some rivers additional difficulty arises from very high water conductivity (e.g., in the Vistula River the conductivity regularly reaches 1500-2000 $\mu\text{S cm}^{-1}$ and often exceeds even 3000 $\mu\text{S cm}^{-1}$) making electrofishing nearly impossible and inefficient. For these reasons various approaches have been developed to deal with these difficulties and to maximize sampling efficiency and precision. To obtain these goals, often specific sampling gears are used or data from multiple gears are used jointly (Casselman et al. 1990, Lapointe et al. 2008, Freedman et al. 2009, Herzog et al. 2009, Szalóky et al. 2014, Zajicek and Wolter 2018) or sampling design is modified to meet specific aspects of the ecology of a certain species of interest (Johnson and Covich 2000, Kocovsky et al. 2010, Nunn et al. 2014).

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One of the biological phenomena that might be incorporated into monitoring design are diel differences in exploitation of certain habitats by particular fish species. This diel changeover of fish assemblages in shallow habitats has been described for rivers of various sizes in different climate zones (Copp 2010 and literature cited therein).

In this paper we describe diel differences in the abundance of three small-bodied, bottom-dwelling fish species listed in the Annex II of the Habitat Directive; thus, they are subjects of regional- and EU-level monitoring. The whitefin gudgeon, *Romanogobio belingi* (Slastenenko, 1934), and the golden loach, *Sabanejewia baltica* (Witkowski, 1994), are listed as vulnerable (VU), whereas Kessler's gudgeon, *Romanogobio kesslerii* (Dybowski, 1862), as near threatened (NT) on the Polish regional red list (Witkowski et al. 2009). All three are considered rare species with small local populations. In recent years numerous new locations of *R. belingi* and, to a lesser extent, of *R. kesslerii* were found throughout Poland, and it was even hypothesized that they might be undergoing expansion (Nowak et al. 2013, 2014). The reverse trend has been observed recently for *S. baltica*, which has declined in or completely disappeared from a number of locations (Kotusz 2017). However, the only sampling technique for all these species approved by national institutions is electrofishing (Makomaska-Juchiewicz and Baran 2012). It is suspected that this method is strongly biased against small-bodied fishes in large rivers, resulting in numerous false-negative records and the severe underestimation of their population parameters. For this reason in the present study we propose an alternative, or rather supplementary, technique for sampling *R. belingi*, *R. kesslerii*, and *S. baltica* in large lowland rivers.

Material and methods

Sampling was performed in a shallow sandy habitat formed between old dykes on the left bank of the Vistula River close to the village of Łęka near Nowy

Korczyn (lat. 50.28272°N, long. 20.84982°E). The river bed in this location is 160–230 m wide and the average discharge is approximately $220 \text{ m}^3 \text{ s}^{-1}$ (during periods with discharge greater than $300 \text{ m}^3 \text{ s}^{-1}$ sampling is impossible because of water depth greater than 1.3 m over the sandbar). Fish were caught using a 15-m-long seine net (two 5×1.5 m wings and a central bag of 5×3 m made of knotless nylon netting, bar mesh size = 8 mm), hauled downstream at approximately 50 m transects. Samples were taken every 3 h over a 24 h period from August 2014 to March 2016 (a total of seven sampling sessions: August 21–22, 2014; November 5–6, 2014; March 9–10, 2015; May 7–8, 2015; July 7–8, 2015; December 17–18, 2015; March 16–17, 2016). After capture, the fish were measured (standard length, SL ± 1 mm) and released back into the water. At each sampling event, illumination at the water surface was measured (± 1 lx, Milwaukee MW700 lux meter), and dusk was considered to be the time when illumination dropped below 0.001 klx.

The data are presented as CPUE (number of specimens caught in a single haul). To obtain 95% confidence intervals for the mean CPUE, we applied the bootstrapping technique (10,000 iterations; Haddon 2011). Diel differences in the CPUE of every species were investigated with generalized linear mixed-effects models (GLMM) with Poisson distribution for error and the log-link function. As 24-h variation was of primary interest, we set season (spring, summer, fall, and winter) as the random effect. The calculations were performed in the R environment (R Core Team 2018) with the *rcompanion* (Mangiafico 2018) and *lme4* packages (Bates et al. 2015).

Results

During the study, a total of 1,812 specimens belonging to 24 fish species were collected. The most abundant species were dace, *Leuciscus leuciscus* (L.), and bleak, *Alburnus alburnus* (L.). They constituted up to 10.8% and 8.9% of the total, respectively. The next

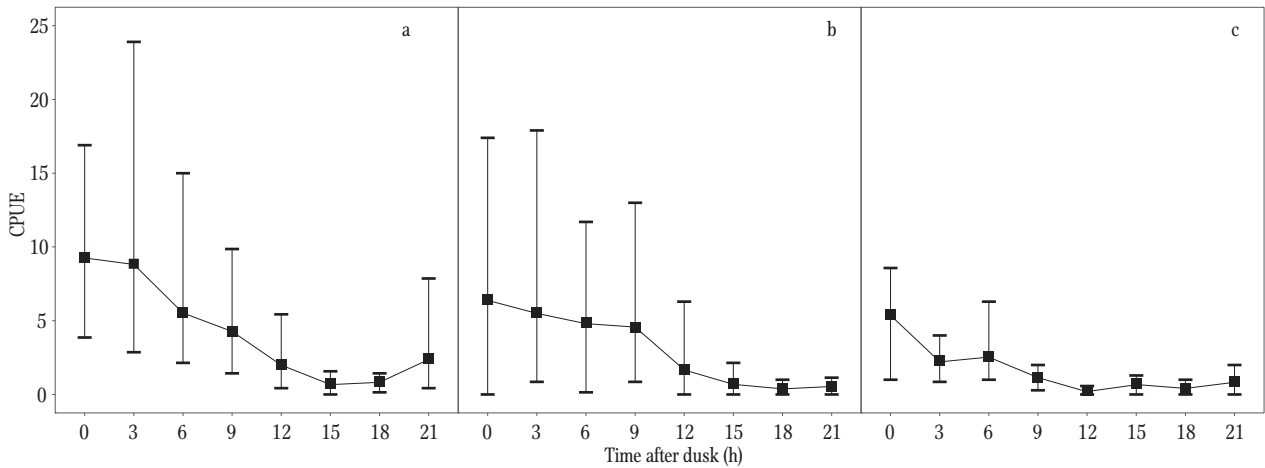


Figure 1. Mean (\pm 95% C.I.) abundance of (a) *R. belingi*, (b) *R. kesslerii*, and (c) *S. baltica* at different time intervals after dusk.

three most abundant species were: *R. belingi* (237 specimens at 5.0% of the total), *R. kesslerii* (174 specimens, 3.6%), and *S. baltica* (96 specimens, 2.0%). The lengths of the individuals collected ranged between 12–93 mm SL for *R. belingi* (mean $46.2 \pm$ S.D. 16.5 mm), 25–76 mm SL for *R. kesslerii* (36.9 ± 7.6 mm), and 21–75 mm SL for *S. baltica* (58.1 ± 7.7 mm). Standard length did not differ among diel periods in any of these three species (GLMM, $P > 0.05$).

A clear diel pattern of abundance was found in all three species investigated (Fig. 1); they were caught in the highest numbers within the first 3 h after dusk. Then their abundance gradually decreased by approximately 0.88 to 0.89 times for every additional h after dusk (GLMM, $z = -9.88$, $P < 0.001$ for *R. belingi*, $z = -8.97$, $P < 0.001$ for *R. kesslerii* and $z = -7.00$, $P < 0.001$ for *S. baltica*). When all the observations were pooled into only two categories, nighttime abundance was 6.62 times higher than that of daytime for *R. belingi* (GLMM, $z = 10.32$, $P < 0.001$), 8.84 times higher for *R. kesslerii* ($z = 8.35$, $P < 0.001$), and 4.30 times higher for *S. baltica* ($z = 6.30$, $P < 0.001$) (Fig. 2).

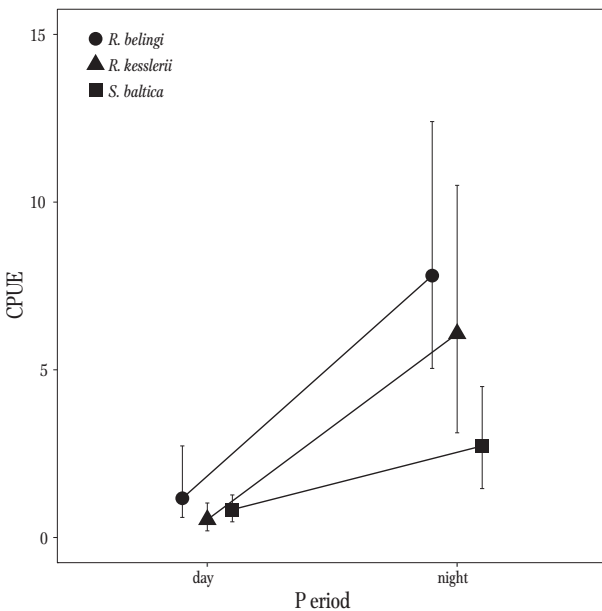


Figure 2. Mean (\pm 95% C.I.) day and night abundance of *R. belingi*, *R. kesslerii*, and *S. baltica*.

Discussion

Romanogobio belingi, *R. kesslerii*, and *S. baltica* are small-bodied, bottom-dwelling fish species, which makes them particularly difficult to sample in large rivers with deep waters and low transparency, especially with electrofishing. These three species constituted up to 10.6% of all the fish collected during the study period. For comparison, only 14 specimens of *R. belingi*, 4 specimens of *R. kesslerii*, and no *S. baltica* were caught at 13 sites (each 1,000 m in length) on the Vistula River between the mouths of the Raba and Wisłoka rivers. Each of these sites was

Table 1

Comparison of beach seining (this study) and electrofishing at corresponding river stretches (spring and fall sampling; Nowak, unpublished data); the species discussed in this study are highlighted in bold

Species	Beach seining	Electrofishing
<i>Leuciscus leuciscus</i>	515	3
<i>Alburnus alburnus</i>	425	439
<i>Romanogobio belingi</i>	238	0
<i>Romanogobio kesslerii</i>	174	0
<i>Sabanejewia baltica</i>	96	0
<i>Gobio gobio</i>	89	5
<i>Squalius cephalus</i>	51	111
<i>Chondrostoma nasus</i>	48	15
<i>Vimba vimba</i>	45	0
<i>Sander lucioperca</i>	37	0
<i>Barbus barbus</i>	30	11
<i>Blicca bjoerkna</i>	16	11
<i>Leuciscus aspius</i>	7	3
<i>Rutilus rutilus</i>	6	4
<i>Pseudorasbora parva</i>	3	0
<i>Alburnoides bipunctatus</i>	2	4
<i>Carassius gibelio</i>	2	0
<i>Gymnocephalus cernua</i>	2	0
<i>Leuciscus idus</i>	2	3
<i>Lota lota</i>	2	1
<i>Perca fluviatilis</i>	2	8
<i>Barbatula barbatula</i>	1	4
<i>Cottus microstomus</i>	1	0
<i>Scardinius erythrophthalmus</i>	1	1
<i>Rhodeus amarus</i>	0	36
<i>Abramis brama</i>	0	9
<i>Silurus glanis</i>	0	7
<i>Cyprinus carpio</i>	0	1
<i>Esox lucius</i>	0	1
Species richness	24	20
Total abundance	1,812*	677

* Including 14 specimens of unidentified cyprinid larvae

electrofished twice, in spring and in fall 2015 (Nowak, unpublished data). Moreover, electrofishing completely failed to record any of these species at the

site where beach seining was performed (Table 1). This suggests that all three species might be seriously overlooked and their populations underestimated during routinely performed electrofishing.

Both of the gudgeon species, *R. belingi* and *R. kesslerii*, and *S. baltica* are considered species of primarily nocturnal activity (Ahnelt and Tiefenbach 1998, Bless and Riehl 2004, Przybylski 2012). Quantitative data on *S. baltica* are scarce, but there is a bulk of evidence that in temperate lowland rivers gudgeon species usually come inshore after dusk (Mathews 1971, Copp and Jurajda 1993, 1999, Koščo et al. 2008, Nowak 2017). Our results clearly show that nighttime samples might be several times (in the present study from 4.3- to 8.8-times, depending on the species) higher than those collected during the daytime. Similar conclusions have already been reached for different species in various habitats (Sanders 1992, Johnson and Covich 2000, Kocovsky et al. 2010, Szalóky et al. 2014, Zajicek and Wolter 2018), but these refer especially to small, bottom-dwelling fishes in large lowland rivers (Wolter and Freyhof 2004, Koščo et al. 2008, Lenhardt et al. 2017, Nowak 2017). However, in smaller water courses daytime samples might be fully representative of actual fish assemblages (Copp 2010 and literature therein, Czeglédi et al. 2016). Taking into account the restricted resources available for any monitoring program, it seems crucial to apply methods and gears that provide the best catch-to-effort ratios. Whereas nighttime electrofishing might not be preferable because of safety considerations, and only daytime electrofishing is approved by the national monitoring methodology (Makomaska-Juchiewicz and Baran 2012), using a relatively small seine net in shallow, unstructured habitats during periods of darkness seems both straightforward and efficient. Beach seining is one of the standard methods for monitoring fishes in North American rivers (Guy et al. 2009) and additional gears are commonly applied worldwide (e.g., Guy et al. 2009, Zajicek and Wolter 2018).

Differences in the day and night abundances of the species studied in this paper might be attributed

to the varied catch efficiency of seine nets at different light levels (Pierce et al. 2001, Říha et al. 2008). On the other hand, clear patterns of diel dynamics were found in all three species. If differences in abundance were to be attributed to different catch efficiency, such a gradual change in the numbers of the fish collected would be rather unexpected. Irrespective, when considering monitoring design it seems less important if the higher abundance of target species at dusk might be attributed to the specific aspects of their biology or to improved gear performance as far as it permits optimizing catch and effort.

The need for adapting monitoring techniques to specific attributes of target species or their habitats is often neglected, which can lead to biased results (Nunn et al. 2014, Lintermans 2015, Stewart et al. 2017). On the other hand, sound data on distribution and abundance are crucial for every management plan of endangered or rare species. The present study demonstrates that for *R. belingi*, *R. kesslerii*, and *S. baltica* beach seining after dusk might be very efficient. For this reason we strongly recommend considering it as a supplementary inventory technique in monitoring programs targeting any of these three species. Obviously, the standardization of this method for the requirements of monitoring is beyond the scope of this paper and should be done based on a much broader spatial scale.

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